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1 Status and perspectives for pikeperch (*Sander lucioperca*) stocks in the Baltic Sea 2 region and central Europe

3

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23

24 Abstract

25

26 Pikeperch (*Sander lucioperca*) is a European fresh and brackish water piscivorous fish, important as a key
27 predator and a valuable fisheries species. Despite concerns that some stocks are depleting due to overfishing
28 and environmental changes, stock assessments are implemented sporadically. We provide an overview of data
29 collection and population assessments currently used for nine pikeperch stocks across six European countries
30 and apply a unified assessment framework (Bayesian surplus production models) to evaluate population status
31 and trends. Our results show that three stocks, including two in the Baltic Sea, are strongly depleted, with
32 estimated biomasses considerably lower than the biomass at maximum sustainable yield (B_{MSY}). Other stocks are
33 close to their estimated B_{MSY} . Further, recent population trends suggest that only one stock (Kvädöfjärden) is
34 increasing, whereas three (Curonian Lagoon, Lipno, Galtfjärden) are rapidly declining. In most cases the stocks
35 with a favourable status or signs of recovery were also those for which strong management strategies have been
36 implemented. Importantly, although most stocks are strongly targeted by recreational fishing, estimates of
37 recreational catch are highly uncertain. We highlight an urgent need to improve pikeperch scientific monitoring
38 and assessment of recreational catches.

39

40 **Keywords:** surplus production models, data limited stocks, inland fisheries, coastal fisheries, recreational fishing

41

1. Introduction

Regular population assessments are more likely to lead to sustainable fisheries (Melnychuk *et al.*, 2021), yet, a majority of the world's fish populations remain unassessed, especially in coastal and inland regions. Pikeperch (*Sander lucioperca*) is a key predatory fish species in European coastal, brackish, and freshwater ecosystems and is highly valued by both commercial and recreational fishers (Arlinghaus and Mehner, 2004). Being one of the largest predators, pikeperch also plays important ecological and regulatory roles (Ranåker *et al.*, 2014). However, many pikeperch stocks are depleted (Mustamäki *et al.*, 2014; FAO, 2022) or their status is uncertain. Globally, reported commercial pikeperch catches declined from 50 thousand tons in 1950 to 14.6 thousand in 2009, although they increased to 26 thousand tons by 2020 (Colchen *et al.*, 2020; FAO, 2022). These statistics may ignore inland fisheries and do not include recreational catches, which for many pikeperch populations likely exceed commercial catches (Heikinheimo *et al.*, 2014; Dainys *et al.*, 2022a, b). Most countries do not collect regular data on recreational catches and their impacts, and given recent advances in recreational fishing technology, there is an urgent need to review and better monitor pikeperch population status.

Despite its importance, current data collection and assessment of European pikeperch populations remains limited. Although some stocks in Finland, Estonia and Sweden have been assessed with age-based virtual population analyses (Eero, 2004; Härkönen *et al.*, 2023; Olin *et al.*, 2023), assessments of most other stocks are based on mean size trends, reported commercial or recreational catches, or are not undertaken at all (see Table S1). Currently, population analyses paint a diverse picture in the Baltic Sea region and central European waters. In Finland's inland and coastal waters, growing commercial and recreational catches suggest improved population status (LUKE, 2022). In contrast, fishery independent monitoring of Swedish coastal pikeperch populations in the northwest region of the Baltic Sea suggests large declines in abundance since 1995 (Mustamäki *et al.*, 2014). A similar decrease in pikeperch abundance is seen for the Pärnu Bay in Estonia (Eschbaum *et al.*, 2021) and the Curonian Lagoon in Lithuania (Andrašūnas *et al.*, 2022), and scientific surveys show that many other populations in the Baltic region have been declining (Bergström *et al.*, 2016). For freshwater populations, declining status of pikeperch stocks led to the complete closure of commercial fishing in Kaunas Water Reservoir in Lithuania (Kaunas) in 2013, but large recreational catches likely precluded the expected population recovery (Dainys *et al.*, 2022a). Similarly, recreational catches are also likely to be the main driver for the population declines in Lipno Reservoir, Czech Republic (Lipno) (de Moraes *et al.*, 2023).

Lack of uniform and regular assessments of pikeperch populations is likely to impede appropriate management. Regular assessments are particularly important considering rapidly changing environmental conditions and the increasing popularity and efficiency of recreational fishing. For example, climate warming and eutrophication are likely to be benefiting some populations, given that pikeperch thrives in warm, turbid, well-oxygenated and productive ecosystems (Veneranta *et al.*, 2011; Müller-Karulis *et al.*, 2013). Similarly, rapid improvement in recreational fishing technology and increasing popularity of catch and release fishing are also likely to be altering mortality and abundance of many populations, although the directions of these changes remain unknown. To address these challenges, this study aims to bring together diverse data sources from a range of European marine and freshwater pikeperch populations and use these data in a standardised manner to conduct a unified analyses of population status. More specifically our aims are: 1) to briefly review current pikeperch stock

83 assessment approaches and management policies across six European countries, 2) apply a unified analysis using
84 Bayesian surplus production modelling (Winker *et al.*, 2018) across all populations using statistically
85 standardised catch per unit effort (CPUE) time series, and 3) provide recommendations for improving data
86 collections and monitoring of pikeperch populations.

87

88 **2. Data and methods**

89

90 **2.1. Study areas and data**

91

92 We analysed nine pikeperch populations from six northern and central European countries (Finland, Sweden,
93 Estonia, Lithuania, Czech Republic and Austria), representing a cross section of marine coastal and inland water
94 bodies with different management strategies (Fig. 1, Table S2). Data used in this study were contributed by the
95 relevant scientists from each country and included >20 years of commercial effort and catches, recreational
96 catches (where available) as well as scientific surveys (Table S3, S4). Further details about the data for each stock
97 are provided in the Supplementary material. During the process of data collection, we also asked experts from
98 each country (i.e., authors of this study) to evaluate the quality of recreational, commercial and scientific survey
99 data on a 1 (unreliable) to 5 (highly reliable) scale (Table S4). Full details on the ranking criteria and data quality
100 assessment results are shown in the Supplementary material.

101

102 **2.2. CPUE estimation and standardisation**

103

104 Depending on data availability, CPUE time series for surplus production models (SPM) were estimated from
105 scientific surveys, commercial or recreational catch and effort data (Table S4). The use of CPUE as an index of
106 relative abundance requires the removal of the effects of variation due to changes in factors other than stock
107 biomass (such as season, location etc.). We standardized CPUE using generalised linear models with a Tweedie
108 distribution (Shono, 2008), as implemented in the R (R Core Team, 2023) packages ‘statmod’ (v. 1.4.33) and
109 ‘tweedie’ (v. 2.3.3). For full details of model selection see Supplementary material, Table S5.

110

111 **2.3. Surplus production models**

112

113 We applied the open-source Bayesian state-space surplus production model (SPM) framework JABBA (Winker *et al.*
114 *et al.*, 2018). The state-space framework enables incorporation of both observational and process errors, and the
115 Bayesian approach allows quantifying uncertainty in the parameter estimates. A critical assumption in SPMs is
116 the percentage of the stock carrying capacity, which gives the highest production and therefore maximum
117 sustainable biomass (B_{MSY}). This can be assumed at 37% if models apply the Fox production curve, 50% for
118 Schaefer’s production curve, or can be set at a higher level with a generalized Pella-Tomlinson (1969) production
119 curve. We chose the Pella-Tomlinson production function and assumed that B_{MSY} occurs at 65% of the stock’s
120 carrying capacity. It is the most conservative approach, which is likely to be more accurate for pikeperch stocks,
121 given the evidence that size structure and stock-specific recruitment conditions can exert large impact on
122 population viability and therefore on MSY (Maunder, 2003; Vainikka and Hyvärinen, 2012; Vainikka *et al.*, 2017).
123 A more conservative and precautionary approach is also warranted given the largely unknown but potentially

124 substantial recreational fishing pressure (e.g., see Embke *et al.*, 2019 for a demonstration of large recreational
125 fishing impact on the North American stocks of a closely related species), as well as lessons learned from overly
126 optimistic model forecasts for other fisheries stocks (Walters and Maguire, 1996). A more commonly used
127 Schaefer function was also applied, with model outcomes provided in the Supplementary material. JABBA
128 requires prior information on the resilience parameter r (intrinsic rate of population increase), carrying capacity
129 K and the relative initial biomass at the beginning of the time series. Priors for K and initial biomass were stock
130 specific, whereas r priors were identical for all stocks and were based on classifying pikeperch as a low resilience
131 species (Froese *et al.*, 2017) (see Supplementary material for further details). To evaluate the model's goodness-
132 of-fit, we used the standard deviation of the normalized residuals (SDNR) and a runs test (Francis, 2011), which
133 assess the randomness and the magnitude of CPUE time series residuals compared to model predictions.

134 135 **2.4. Mean size estimation**

136
137 Given that SPMs do not consider population size structure, we have also explored trends in population mean
138 size across the studied periods. Individual size data were collected either from commercial catches (i.e., all fish
139 above the minimum size limit) or scientific monitoring (the full range of mesh sizes), so that although mean sizes
140 cannot be compared across different sampling methods, they are nevertheless useful as an indication of trends
141 in population body size. To account for variability across sampling seasons, mesh sizes and sites within a
142 population, yearly changes in mean length were estimated using linear models, similar to the CPUE analyses
143 above. Fixed effects, as well as initial and final models varied for each stock, depending on the sampling
144 parameters (Table S5).

145 146 **3. Results**

147 148 **3.1. Data collection and quality and current stock management**

149
150 Based on expert judgement, reporting of commercial catch data was of relatively high quality (data quality
151 values typically ranging at 4-5, Table S4). International cross-jurisdictionally managed Curonian Lagoon and Lake
152 Peipsi had the lowest commercial catch data scores. Commercial effort data were considered of high quality
153 only for the Finnish Archipelago and Lake Oulujärvi. In other places, commercial effort is collected as the total
154 allowable, rather than the actual effort (e.g., number of fishery permits in Lake Constance); is available only on a
155 much larger spatial scale (entire Swedish coast, rather than specific stocks); or reporting requirements have
156 changed through time (Curonian Lagoon) making comparisons unfeasible. Fishery independent scientific surveys
157 are conducted annually in five out of nine stocks with consistent methodology and provide high quality data
158 (scores of 5) (Table S4). For the Finnish Archipelago and Lake Oulujärvi no fishery independent scientific surveys
159 are taking place.

160
161 The main gaps in data availability were in recreational catch and effort. Only five stocks had some estimates of
162 recreational catches, and only one of them (Lipno) was considered of high quality and was based on mandatory
163 catch reports. For other stocks, recreational catch estimates were either limited to short time periods (Kaunas),
164 were based on questionnaires and therefore potentially biased (Finland) or conducted only every few years

165 (e.g., every 5th year in Lake Oulujärvi). In the Baltic Sea stocks (Curonian Lagoon, Galtfjärden, Kvädöfjärden) no
166 assessment of recreational catch was available, although rough estimates along the Swedish coast suggest that
167 recreational catches could exceed the commercial catch (Sundelöf *et al.*, 2022).

168

169 In most European pikeperch stocks studied here, current population assessment methods rely on trends in
170 commercial catches, commercial and scientific CPUE or body size statistics (Table S1). Only two populations
171 (Finnish Archipelago and Lake Oulujärvi) have been assessed with a more complex virtual population analysis
172 (Table S1).

173

174 **3.2. Management policies**

175

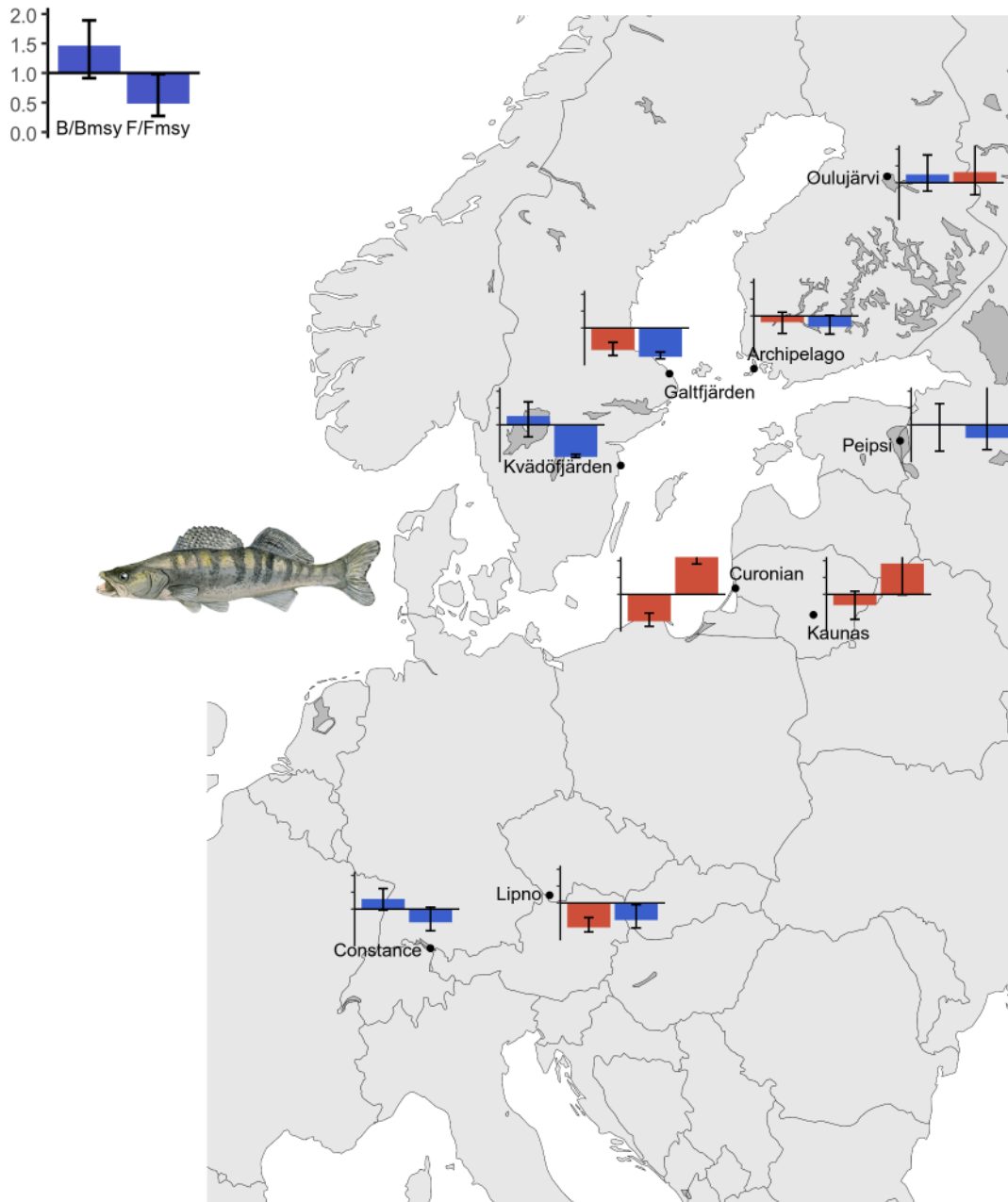
176 In most cases, management policies used for commercial pikeperch harvest are based on either total allowable
177 catch (TAC) (two out of seven commercially harvested populations) or total allowable effort (TAE, setting
178 number and types of gears) for the remaining populations. All populations have size restrictions on harvest, such
179 as minimum mesh size and a minimum landing size, typically set at about 45 cm total length (Table S2). In
180 Curonian Lagoon, TAC for the commercial fishery is implemented on the Russian side, while in the Lithuanian
181 side, management since 2013 applies TAE restrictions only. Seasonal and spatial closures for commercial
182 fisheries apply only in the Curonian Lagoon. For two inland populations (Kaunas and Lipno) commercial fishing is
183 currently banned (since 2013 and 1976 respectively). Recreational fishery management typically involves daily
184 bag limits, minimum landing size (ranging from 40 to 46 cm) and in some cases restrictions on lures (only in
185 Lipno) or seasonal closures lasting from two to six months (Table S2). Total recreational catch is not restricted in
186 any populations. Since 2009, permanent spatial closures have been implemented in Lipno and now encompass
187 6% of the total reservoir area, being the only pikeperch population with permanent spatial closures.

188

189 **3.3. Pikeperch stock status**

190

191 Both Pella-Tomlinson and Schaefer production models produced similar population trends for all stocks in this
192 study. Models using the Pella-Tomlinson production function showed that during 2019–2022 three out of nine
193 studied stocks (Curonian Lagoon, Lipno and Galtfjärden) were confidently in depleted state with the estimated
194 stock biomass and its uncertainty ranges below the B_{MSY} (Fig. 2, for Kobe plots see Fig. S1) and same results were
195 obtained with the Schaefer production model (Fig. S2). Trends in mean length of pikeperch also showed
196 declining trends in the Curonian Lagoon and Lipno but were relatively stable size for Galtfjärden. None of the
197 stocks were confidently above B_{MSY} (Figs 1, 2) using the Pella-Tomlinson model, but four stocks were above B_{MSY}
198 when the Schaefer model was applied. Regardless of the model, four stocks (Kvädöfjärden, Lake Oulujärvi, Lake
199 Peipsi, and Upper Lake Constance) showed relatively good status with biomass levels close to the estimated B_{MSY}
200 and mean sizes for the first three populations also remaining stable. Finally, for Kaunas and Finnish Archipelago
201 the estimated biomasses were slightly lower B_{MSY} , but there was a small increase in mean size in Kaunas and a
202 stable mean size in commercial catches in Finnish Archipelago. In all six populations (Kvädöfjärden, Lake
203 Oulujärvi, Lake Peipsi, Upper Lake Constance, Kaunas and Lipno) the uncertainty ranges overlapped with B_{MSY}
204 values.

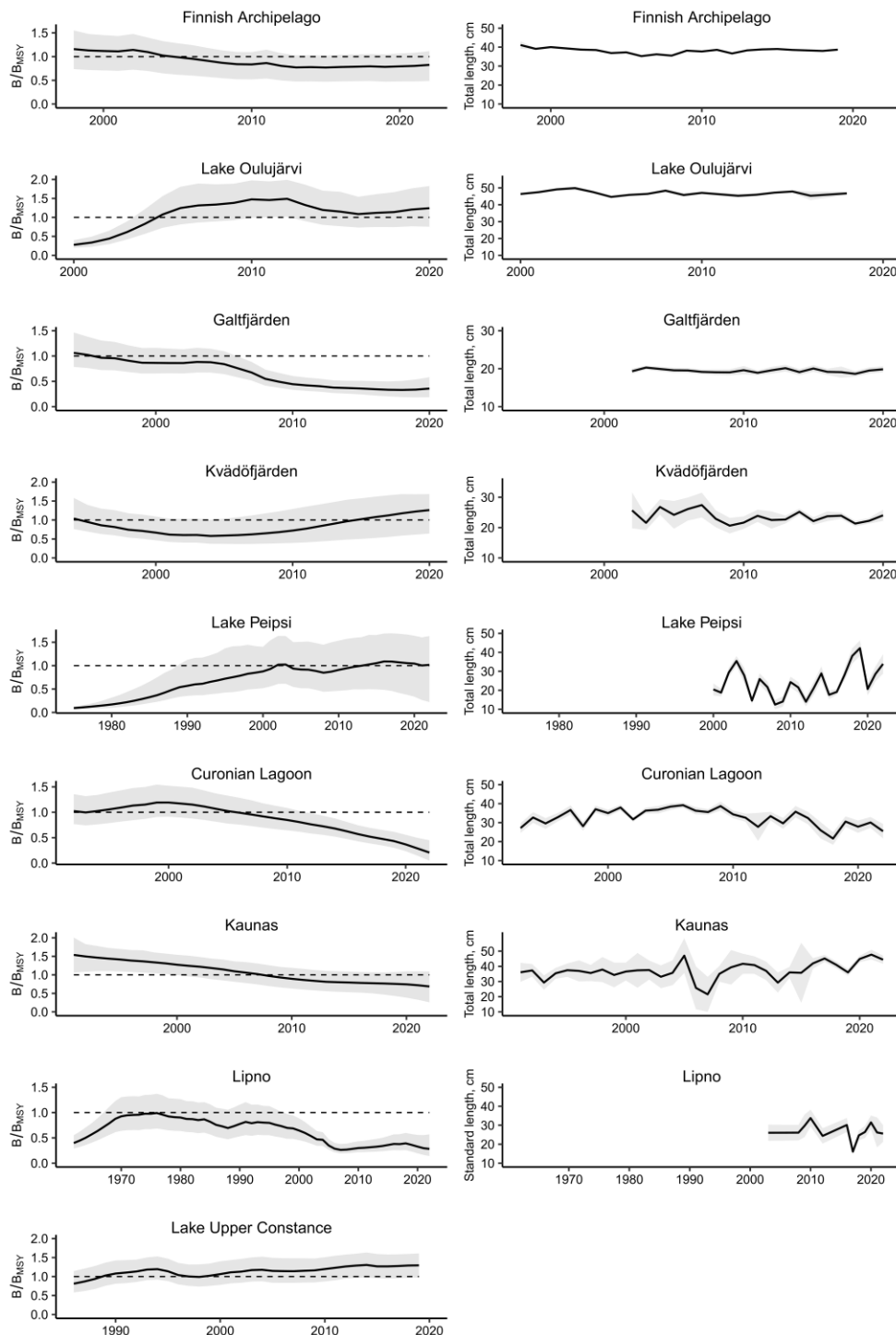


205
 206 **Fig. 1.** Status of pikeperch stocks (last year in the time series, 2019 to 2022) in the European populations
 207 studied. The first bar illustrates the ratio of current biomass (B) versus biomass required for maximum
 208 sustainable yield (B_{MSY}), where values in blue (>1) indicate a good status and values in red (<1) show bad status.
 209 The second bar shows the current fishing mortality (F) versus mortality leading to a maximum sustainable yield
 210 (F_{MSY}) and values <1 indicate good status. Error bars show uncertainty intervals from the surplus production
 211 models (see Fig. 2).
 212

213 When looking at population trends, the status of four pikeperch stocks has improved over the last decades.
214 Specifically, Lake Peipsi pikeperch biomass was low in the 1960–1970s, but since 1990s–2000s the stock is in a
215 good state and still improving (Fig. 2). In Lake Oulujärvi pikeperch population collapsed in the 1960–1970s, and
216 despite stocking since 1985, biomass remained very low until the onset of natural reproduction in 1998.
217 Currently the stock seems to fluctuate around B_{MSY} . The Kvädöfjärden stock along the Swedish coast also
218 showed signs of recovery in biomass, yet mean fish sizes have been decreasing over the last two decades (Fig.
219 2). The Finnish Archipelago population biomass declined in 2000s but appears to have stabilised since 2010.

220

221 In contrast, four other stocks - Curonian Lagoon and Galtfjärden in the Baltic Sea, as well as Kaunas and Lipno –
222 showed a declining biomass over recent decades, with the lowest levels in the Curonian Lagoon and Galtfjärden.
223 Despite declining overall biomass, the mean size of pikeperch in Kaunas has been increasing, whereas in Lipno
224 neither biomass nor mean size has shown any signs of recovery. Finally, the population status in Lake Constance
225 appeared to be good, with biomass above B_{MSY} and relatively low fishing mortality. However, this result is
226 uncertain, because the time series of CPUE and catch data didn't include sufficient contrast in population
227 abundance to estimate population parameters and available data quality was relatively low; as a result, the
228 overall model fit is relatively poor.



229
 230 **Fig. 2.** On the left: Relative pikeperch biomass of the nine stocks assessed. The y axis indicates the biomass
 231 relative to the biomass at maximum sustainable yield (B_{MSY}). The grey area and the thick central line show the
 232 95% and 50% posterior probabilities. Note the different time series for each stock. On the right: Trends in mean
 233 length using data from scientific surveys (Lake Peipsi, Galtfjärden, Kvädöfjärden, Curonian Lagoon, Kaunas,
 234 Lipno) or commercial catches (Finnish Archipelago, Lake Oulujärvi).
 235

236 4. Discussion

237
238 The pikeperch stocks analysed in this study provide a good representation of northern, eastern and central
239 European populations across different countries, habitats, management and monitoring practices and fishing
240 impacts. Some populations are mostly exploited by commercial fisheries (in the Baltic Sea) while others are
241 fished only recreationally (Kaunas and Lipno). Overall, our results show a mixed picture of both declining and
242 improving trends in different areas, with about half of the population being close to their maximum sustainable
243 yield biomass levels and the remaining half being moderately or strongly depleted. However, none of the stocks
244 were confidently above B_{MSY} , while three stocks were confidently below, indicating the need for better
245 management plan for commercially and recreationally intensively harvested pikeperch in Europe.

247 4.1. Trends and status of Baltic Sea populations

248
249 Our study suggests that the most concerning situation occurs in two coastal stocks in the Baltic Sea - Curonian
250 Lagoon and Galtfjärden, the latter finding also supported by decreasing pikeperch age in commercial catches
251 (Sundelöf *et al.*, 2022). In the coastal areas of the Baltic Sea pikeperch traditionally supported important
252 commercial fisheries (Fig. S3), but in recent years its significance has decreased, and sustainability of these
253 fisheries have been questioned (Eero, 2004; Eesti Mereinstituut, 2016). Many fisheries harvest a large
254 proportion of the pikeperch stock soon after or even before reaching maturity, and despite minimum size limits,
255 discard mortality from commercial fisheries can be high (Ložys *et al.*, 2022). Recreational fishing, although often
256 not assessed, is also likely to be intense (Sundelöf *et al.*, 2022), and some stocks may also be facing high
257 predation pressure from seals and cormorants (Mustamäki *et al.*, 2014). One of the most depleted and still
258 declining stock is in the Curonian Lagoon, even though commercial catches over the last decade have remained
259 surprisingly stable (Fig. S3; Andrašūnas *et al.*, 2022). This stability is likely caused by increases in fisheries
260 efficiency through recent adoption of modified fyke nets, which unfortunately also leads to high incidental
261 mortality of undersized fish (Ložys *et al.*, 2022). Further, although not included in this study due to low data
262 availability and quality, the status of the Pärnu Bay (Estonia) pikeperch stock is also considered poor (Eschbaum
263 *et al.*, 2021). This population once yielded substantial commercial catches (800 tons in the 1930s, 150-300 tons
264 in the 1990s), but has collapsed in the 2000s and has since remained in a poor state.

265
266 In contrast to the poor status of Galtfjärden stock along the Swedish coast, the biomass of a nearby
267 Kvädöfjärden (Fig. 2) stock is increasing, and this is also supported by other ecological time series data (HELCOM,
268 2023). However, Kvädöfjärden pikeperch body sizes are decreasing and the reasons for biomass recovery remain
269 unclear. One Baltic Sea area where status of pikeperch stocks appears to be relatively stable and likely improving
270 during recent years is in the Finnish Archipelago. This might be thanks to the increased minimum landing sizes
271 (from 37 cm to 40 or 42 cm TL in 2019 depending on the area, Vainikka *et al.* 2017; Olin and Raitaniemi, 2021)
272 and reductions in commercial fishing effort since 2003 (LUKE, 2022). Increasing water temperature (Viitasalo
273 and Bonsdorff, 2022) and eutrophication may also be improving pikeperch reproduction success and individual
274 growth rates, given that these populations are close to their northern distribution limit. Nevertheless, the
275 recovery of pikeperch in the Finnish Archipelago remains slow, possibly due to bycatch mortality of undersized
276 pikeperch (Olin and Raitaniemi, 2021), cormorant and seal predation (Heikinheimo *et al.*, 2015; but see Olin *et*

277 *al.*, 2023) or recreational fishing. Angling effort in the Finnish Archipelago is high and recreational catches in
278 Finland generally exceed the commercial catch (LUKE, 2022). In fact, most coastal Baltic Sea stocks attract
279 intensive recreational fishing, although the effort and catches remain highly uncertain. For example, in 2014
280 recreational catches along the coasts of Sweden were estimated to be in the range of 9 to 64 tonnes compared
281 to just under 14 tonnes in the commercial fishery (Sundelöf *et al.*, 2022). Although uncertain, estimates for
282 recreational catches in the Pärnu Bay and Pärnu river in 2022 were around 26 tonnes, compared to 22 tonnes
283 for the commercial fishery (Kantar, 2023). Unfortunately, for the Curonian Lagoon, even an approximate
284 estimate is not available.

285

286 **4.2. Origins and status of freshwater populations**

287

288 The natural distribution of pikeperch in Europe is largely confined to coastal and brackish habitats and river
289 deltas. Many freshwater populations in Finland, Czech Republic, Lithuania and Austria have been established by
290 stocking, although many now also have natural recruitment (Table S2). In Lipno and Kaunas pikeperch occurred
291 naturally before the respective damming of the Vltava and Nemunas rivers (Vostradovsky and Tichy, 1999), but
292 populations were further augmented by stocking. Several important native freshwater pikeperch stocks in
293 Finland, including the Lake Oulujärvi population, collapsed during the 1960s and early 1970s potentially due to
294 overfishing and low temperatures at the time (Colby and Lehtonen, 1994), and have been restored through
295 reintroductions and subsequent stocking. Currently, pikeperch stock status in Oulujärvi is reasonably good, and
296 this is also supported by an age-structured model applied for this stock (Harkonen *et al.*, 2023). In large lakes,
297 such as Oulujärvi in Finland (ca 900 km²) or Lake Peipsi in Estonia (ca 3550 km²) commercial fishing seems to be
298 considerably more important than recreational catches (Kantar, 2023) and hence harvest controls may be easier
299 to impose. Both Oulujärvi and Peipsi stocks are in an overall good status, with population biomasses close the
300 B_{MSY} limits and mean sizes of fish remaining stable or increasing.

301

302 The situation is different in Lipno and Kaunas reservoirs, which are among the most popular national angling
303 destinations, are considerably smaller (ca 50-60 km²) and therefore easily accessible by anglers. Despite
304 commercial fishing bans in both reservoirs, pikeperch stocks there remain below B_{MSY} limits. There is evidence
305 for some recovery in Kaunas with mean fish body sizes increasing slightly, but such recovery does not seem to
306 occur in Lipno. This lack of recovery is perhaps not surprising, given a rapid increase in recreational angling
307 efficiency (Cooke *et al.*, 2021) and the fact that the relative role of commercial catches might have been small
308 anyway. For example, assessment of recreational effort and catches showed that recreational catches in Kaunas
309 reservoir exceed former commercial catches several fold (Dainys *et al.*, 2022a, b). Recreational fishing also
310 creates an indirect negative effect on pikeperch populations, by increasing competition between pikeperch
311 juveniles and cyprinid fish species. As proportions of predatory fish in populations decline due to intensive
312 harvesting (Myers *et al.*, 2003; Fogliarini *et al.*, 2021), cyprinids, such as bream and roach, often become the
313 dominant species in freshwater and brackish coastal systems (Dainys *et al.*, 2022a). Both young pikeperch and
314 cyprinids compete for food in their early life, yet under low food conditions cyprinids can switch to feeding on
315 detritus and periphyton (Richeux *et al.*, 1992), whereas pikeperch depend entirely on animal food.

316 Despite the importance of recreational catches on freshwater pikeperch populations, regular recreational fishing
317 assessments are largely absent in most countries, except for the Czech Republic, where recreational catch
318 reporting has been mandatory for decades. Recreational angling management remains largely restricted to

319 minimum size and daily bag limits, without attempting to limit the overall catch. A noticeable exception is Lipno,
320 where a permanent spatial closure has been implemented since 2009, comprising 6% of the total reservoir area.
321 Pikeperch was found to be more abundant and larger sized in the protected areas (de Moraes *et al.*, 2023). The
322 main management measure incurring significant expenditure applied in most countries remains regular
323 restocking of pikeperch. This restocking occurs despite consistent evidence that it may contribute little to the
324 replenishment of naturally reproducing populations and may have negative impacts due to introduction of
325 diseases, parasites and mixing of genetic lineages (Radinger *et al.*, 2023). For example, practically all pikeperch
326 stocking programmes in Finland have relied upon only a few source stocks, all of them of southern origin
327 (Salminen *et al.*, 2012). The contemporary pikeperch stock in Lake Oulujärvi (64°N) stems from the Lake
328 Vanajavesi (61°N) stock (Salminen *et al.*, 2012) and despite effective natural reproduction in Lake Oulujärvi since
329 1998 (Vainikka and Hyvärinen, 2012), the population is still annually enhanced by releasing about 300,000
330 young-of-the-year fish. Regular stocking is still applied in Lipno, despite evidence that only about 20% of adult
331 fish derive from stocking (Souza *et al.*, unpublished data). A similar situation occurs in Lake Constance, where
332 pikeperch were originally introduced in the late 1800s and are stocked almost annually, despite studies showing
333 that stocking has limited benefits (Löffler, 1998).

334 **4.3. Future research needs and perspectives**

335
336 Given that the majority of freshwater and coastal fish populations do not feature regular ageing data and
337 therefore cannot be assessed using age-structured stock assessment models (Costello *et al.*, 2012), data-limited
338 methods such as SPM can provide an important alternative approach. In our study, results of SPM for Lake
339 Oulujärvi concurred with outputs from an age-structured model for the same population (Harkonen *et al.*,
340 2023). Despite their simplicity, SPMs are widely used for stock assessment (Wang *et al.*, 2014; Cousido-Rocha *et al.*,
341 2022) and are among the simplest of full stock assessment tools which can inform managers about reference
342 points, stock size and recommended harvest rates (Ludwig and Walters, 1985, 1989). However, SPMs make a
343 key assumption that population abundance is driven by catches, population regeneration or growth rate (r) and
344 carrying capacity (K). This means that all aspects of production are pooled into one production function ignoring
345 age, size, and sex structure. Environmental variation is also either ignored or incorporated into process
346 uncertainty, as is the case with JABBA applied here. SPMs perform best when time series of catch and
347 population abundance data have sufficient contrast to estimate population growth parameters. If data lack
348 contrast, this may suggest that catches have always been managed at sustainable levels or that the stock may be
349 more influenced by factors other than catch e.g., environmental variation (Haddon, 2020). This is likely to be the
350 case for Upper Lake Constance, where pikeperch is not an important target species, and the available data
351 provide little information regarding stock status.

352
353 An important next step to improve assessments and management of pikeperch populations is regular collection
354 of fish body size data. Another essential step would be to focus on assessing and managing recreational catches.
355 Despite growing global recognition that recreational fishing may have large impacts on populations, most
356 countries do not require anglers to report their catches (Cooke *et al.*, 2015, Potts *et al.*, 2020). Although
357 proportions of fish being released during angling trips may be increasing, catch and release angling still
358 contributes to population mortality (Arlinghaus *et al.*, 2002; Sass and Shaw, 2020). Maximum size limits to create
359 harvest slots are only implemented for coastal fish populations in Sweden, despite evidence that larger and

360 older females of many fish species have better spawning success than younger ones (Berkeley *et al.*, 2004; Olin
361 *et al.*, 2012). Permanent spatial closures that ban all fishing are still very rare in the countries included in this
362 study and are not imminent, despite their demonstrated importance for population resilience and protection of
363 important species traits, often selectively removed by anglers (de Moraes *et al.*, 2023). Such strategically located
364 spatial closures could help build resilience against impacts of climate change and environmental fluctuations
365 (Olin and Raitaniemi, 2021). They can also protect a fraction of a population from human induced evolutionary
366 trends, which could have large impacts on pikeperch stock dynamics (Vainikka and Hyvärinen, 2012; Vainikka *et*
367 *al.*, 2017).

368
369 Climate warming and eutrophication of coastal and inland waters are also likely to have some positive effects on
370 pikeperch stocks. In general, pikeperch are believed to benefit from eutrophication and rising water
371 temperatures, because turbid, warm water positively influences pikeperch spawning success and increases
372 somatic growth rates (Pekcan-Hekim *et al.*, 2011; Veneranta *et al.*, 2011). This is especially true for populations
373 in Northern Europe, where pikeperch recruitment and growth are correlated with summer temperatures
374 (Tesfaye *et al.*, 2023). With rapid climate and other environmental changes coupled with increasing popularity of
375 recreational fishing, it is important that pikeperch populations are regularly monitored with consistent and
376 comparable scientific methods and that the scientific community further improves its data sharing and
377 collaboration on population status assessments.

378

379 **CRedit authorship contribution statement**

380

381 AA, HG, EJ: Conceptualization, methodology. EJ, AA, AŠ: Formal analysis. AA: Supervision. EJ, LL, LSH, PH, RN, JK:
382 Data curation. AA, LL: Funding acquisition. EJ, AA: Visualization. EJ, HG, AA, MO, JTD, RN: Writing – original draft.
383 All: Investigation, Writing – review & editing.

384

385 **Data availability**

386

387 Raw data from scientific, commercial and recreational catches and effort can be requested from respective
388 institutions and authors. Standardised CPUE effort and catch times series used in this analysis, together with R
389 codes are available on <https://github.com/eglejak/pikeperchEurope> and archived on Zenodo (DOI:
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391

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403

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