



# Water dispersal of methanotrophic bacteria maintains functional methane oxidation in *Sphagnum* mosses

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It is known that *Sphagnum* associated methanotrophy (SAM) changes in relation to the peatland water table (WT) level. After drought, rising WT is able to reactivate SAM. We aimed to reveal whether this reactivation is due to activation of indigenous methane (CH<sub>4</sub>) oxidizing bacteria (MOB) already present in the mosses or to MOB present in water. This was tested through two approaches: in a transplantation experiment, *Sphagnum* lacking SAM activity were transplanted into flark water next to *Sphagnum* oxidizing CH<sub>4</sub>. Already after 3 days, most of the transplants showed CH<sub>4</sub> oxidation activity. Microarray showed that the MOB community compositions of the transplants and the original active mosses had become more similar within 28 days thus indicating MOB movement through water between mosses. *Methylocystis*-related type II MOB dominated the community. In a following experiment, SAM inactive mosses were bathed overnight in non-sterile and sterile-filtered SAM active site flark water. Only mosses bathed with non-sterile flark water became SAM active, which was also shown by the *pmoA* copy number increase of over 60 times. Thus, it was evident that MOB present in the water can colonize *Sphagnum* mosses. This colonization could act as a resilience mechanism for peatland CH<sub>4</sub> dynamics by allowing the re-emergence of CH<sub>4</sub> oxidation activity in *Sphagnum*.

**Keywords:** plant–microbe interaction, peatland, *pmoA*, microarray, qPCR, ecosystem resilience

## INTRODUCTION

Peatlands store over one third of global terrestrial carbon (Gorham, 1991). Although these ecosystems are carbon dioxide (CO<sub>2</sub>) sinks they are also a major source of methane (CH<sub>4</sub>), formed as the final product of anaerobic degradation of organic matter. Most carbon in these systems is derived from *Sphagnum* mosses (Clymo and Hayward, 1982), the dominant plant in bog-type northern peatlands. Mosses sequester atmospheric CO<sub>2</sub> directly through photosynthesis. Methanotrophic bacteria (MOB) living inside the moss hyaline cells and on leaf surfaces (Raghoebarsing et al., 2005; Kip et al., 2010) also play an important role in carbon binding. These bacteria provide CO<sub>2</sub> for the plant via CH<sub>4</sub> oxidation, a mechanism that is especially important in submerged conditions where CO<sub>2</sub> diffusion is slow (Kip et al., 2010). This phenomenon is of local and global importance as it has been detected in all 23 *Sphagnum* species of a peatland area (Larmola et al. (2010)) and in geographically distant peatlands (Kip et al., 2010) and may be partly responsible for the lower CH<sub>4</sub> emissions of *Sphagnum* bogs in relation to other peatland types (Nykänen et al., 1998).

About 10–15 (Raghoebarsing et al., 2005) or 10–30% (Larmola et al., 2010) of *Sphagnum* biomass carbon is from CH<sub>4</sub> oxidation by MOB. Thus, it seems clear that mosses benefit from their

partners and the relationship has been discussed to be symbiotic (Raghoebarsing et al., 2005). Still, there is evidence that the bacteria involved are only loosely connected to *Sphagnum* (Basiliko et al., 2004; Larmola et al., 2010). The study by Larmola et al. (2010) showed that peatland water table (WT) level is the main factor influencing MOB activity in mosses. *Sphagnum* associated methanotrophy (SAM) became de-/reactivated upon natural WT fluctuation. However, Larmola et al. (2010) did not provide evidence whether reactivated CH<sub>4</sub> oxidation was caused by reactivation of the original MOB community, invasion of new MOB from the surrounding water or by both mechanisms. The ability of MOB to colonize *Sphagnum* from surrounding water would make ecosystem CH<sub>4</sub> dynamics less vulnerable to extended periods of drought than a tight symbiosis between MOB and *Sphagnum* or relying on the reactivation of original community. To test the importance of colonization we examined the question more thoroughly. First, we conducted a similar transplantation trial as in Larmola et al. (2010) where inactive mosses were planted next to active ones. Colonization process was followed by measurement of CH<sub>4</sub> oxidation potentials and community analysis by a microarray that profiles diversity within the *pmoA* gene coding for particulate methane mono-oxygenase (pMMO), a key enzyme in CH<sub>4</sub> oxidation (Bodrossy et al., 2003). By using this method covering

a wider range of MOB diversity than Larmola et al. (2010) we aimed to reveal more detailed changes in community compositions. We hypothesized that colonization of MOB through the water phase is a substantial reason for methanotrophic reactivation. Since we presume that all mosses are colonized through the same pathway this should be reflected in MOB of the neighboring mosses influencing the microbial community of the transplanted moss. Second, we tested the hypothesis in the laboratory by treating inactive *Sphagnum* mosses with water from a wet depression (flark) harboring methanotrophic active *Sphagnum* mosses. As a control, parallel samples were treated with the same water after MOB removal through filtration. CH<sub>4</sub> oxidation potentials were measured and MOB communities analyzed by *pmoA*-based quantitative PCR (qPCR) and denaturing gradient gel electrophoresis (DGGE) analysis followed by sequencing.

## EXPERIMENTAL PROCEDURES

### TRANSPLANTATION EXPERIMENT

#### *Sphagnum* transplantation

The experiment was conducted at the Lakkasuo mire (61°48'N, 24°19'E; 150 m. a.s.l.), a boreal raised bog complex in southern Finland. On 7 July 2008, patches (8 cm in diameter) of inactive *Sphagnum rubellum* from site O were transplanted to six different flark sites (A–F) showing high *Sphagnum* associated methanotrophic (SAM; CH<sub>4</sub> oxidation) activity (Figure 1). To control the effect of transplantation, *S. rubellum* was replanted in the original site and the native *Sphagnum* species (Table A1 in Appendix) gathered from each of six different flark sites were also returned to their original places. Thus, all samples were of transplanted *Sphagnum*. Moss samples were gathered at the beginning of the experiment (0 day), after 3 days, and 28 days. After gathering, mosses were rinsed with deionized water and dried overnight at +4°C. Only

upper 10 cm of the moss plants were included in the following analysis. Ecological variables of the transplantation sites are listed in Table A1 in Appendix.

#### Methane oxidation potential

Methane oxidation potentials were measured as described in Larmola et al. (2010). Briefly, 30 g of moss was incubated in a 600 mL flask with an initial CH<sub>4</sub> concentration of 10000 ppm in the dark at +15°C and the oxidation was monitored after 24 and 48 h by gas chromatography. Results are presented in micromole CH<sub>4</sub> per gram dry weight per hour ( $\mu\text{mol g dw}^{-1} \text{h}^{-1}$ ).

#### Analysis of methanotrophic community composition by *pmoA*-microarray

Community composition of MOB in *Sphagnum* samples was investigated using a microarray (Bodrossy et al., 2003) designed to detect diversity within the *pmoA* gene. DNA was isolated as in Siljanen et al. (2011). A fragment of the *pmoA* gene was amplified using a semi-nested PCR approach with primer pairs A189f/T7–A682r and A189f/T7–mb661r as in Siljanen et al. (2011) with the exception that after the first PCR-step, products not detectable on the gel were diluted 1:10 before being used as templates in the second PCR. Concentration of PCR products was quantified using a Qubit fluorometer (Invitrogen, Carlsbad, CA, USA). *In vitro* transcription and hybridization was performed as in Stralis-Pavese et al. (2004) and the applied probe set was similar to that applied by Abell et al. (2009). One of three parallel original inactive moss samples from the 0 day time point could be successfully analyzed (see Larmola et al. (2010) that SAM inactive *Sphagnum* mosses host MOB DNA). Probing *pmoA* diversity cannot detect *Methylocella* or the recently discovered *Methyloferula* (Vorobev et al., 2011) methanotrophs as the pMMO enzyme is not present in these bacteria. It should be noted that another newly discovered methanotroph group, *Verrucomicrobia*, is also not detectable by the probes we used.

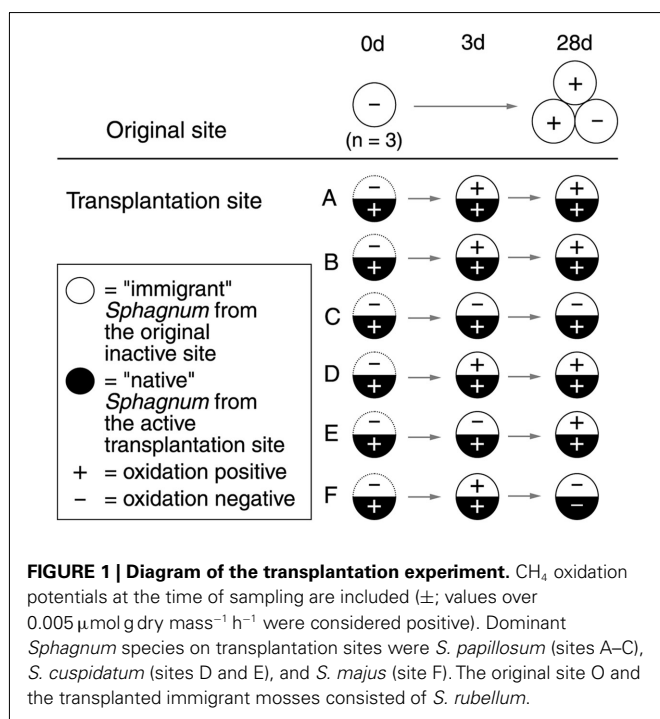
#### Statistical analysis of microarray data

The quantitative nature of the microarray data was converted to a binary matrix (presence = 1, absence = 0) to reveal the community changes caused by different groups of MOB colonizing mosses after transplantation, and also to prevent false interpretation originating from non-quantitative nested PCR approach. The data were then analyzed using principal component analysis (PCA) carried out with CANOCO Version 4.52 (ter Braak and Smilauer, 2002). Sites A–F (Figure 1) were analyzed separately ( $n = 1$ ) and as parallel samples ( $n = 6$ ). Universal MOB probes (positive controls) and probes not hybridizing to any of the samples were excluded from the analyses (threshold for positive samples  $\geq 3$  after normalization of the data to the scale of 0–100).

### BATHING EXPERIMENT

#### Flark water bathing

*Sphagnum* mosses were gathered from Sallie's Fen in NH, USA (43°12.5'N, 71°03.5'W 110 m. a.s.l.). Triplicate ( $n = 3$ ) fresh samples of approximately 30 mL (volume based on the volume of water replaced by the mosses) of inactive *S. magellanicum* were subjected to the following treatments: (I) no treatment; (II) overnight (11 h)



incubation in SAM active *S. majus* 200 mL flark; (III) overnight incubation in 200 mL 0.45  $\mu\text{m}$  filtered SAM active *S. majus* flark water; (IV) overnight incubation in 200 mL SAM active *S. majus* flark water followed by rinsing with deionized water. In the final treatment, *S. majus* gathered from the active (flark) site was included as a positive control ( $n = 3$ ) in the analyses described below. SAM active flark water was collected directly from a wet depression next to *S. majus* vegetation and did not contain any macroscopic plant material. Each overnight incubation was conducted in the dark at  $+20^\circ\text{C}$ . Following treatment, all mosses were dried overnight at  $+4^\circ\text{C}$ . Only upper 10 cm of the moss plants were included in the following analysis.

### Methane oxidation potential and statistical analysis

Methane oxidation potentials were measured as above in the transplantation experiment but on a Shimadzu 14A gas chromatograph equipped with a flame ionization detector (Shimadzu Corp., Kyoto, Japan). The results are presented in micromole  $\text{CH}_4$  per gram dry weight per hour. The difference between sample treatments was tested using Kruskal–Wallis non-parametric analysis of variance followed by Nemenyi test for pairwise comparisons ( $p < 0.05$ ; Zar, 1999).

### Analysis of community composition by DGGE and sequencing

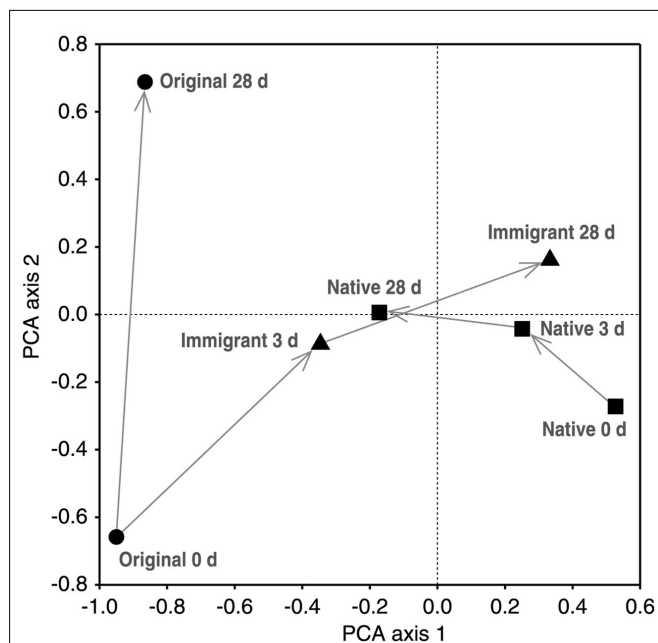
Diversity of MOB in *Sphagnum* samples from the bathing experiment was explored by *pmoA*-based PCR–DGGE analysis and sequencing as previously described (Tuomivirta et al., 2009; Larmola et al., 2010) using the primer pair A189f/GC-621r designed to target methanotrophs abundant in boreal peatlands (Tuomivirta et al., 2009). DNA was isolated as above in the transplantation experiment. Determined *pmoA* gene sequences were submitted to Genbank under accession numbers HQ651182 and HQ651183.

### Quantification of *pmoA* genes and statistical analysis

Quantitative PCR was carried out as previously described (Tuomivirta et al., 2009) using the same primer pair as in the DGGE analysis (A189f/GC-621r). Results are expressed in *pmoA* copy number per gram dry weight. To test the difference between sample treatments, values were  $\ln$  transformed to normalize the data followed by ANOVA and Tukey's HSD test ( $p < 0.05$ ).

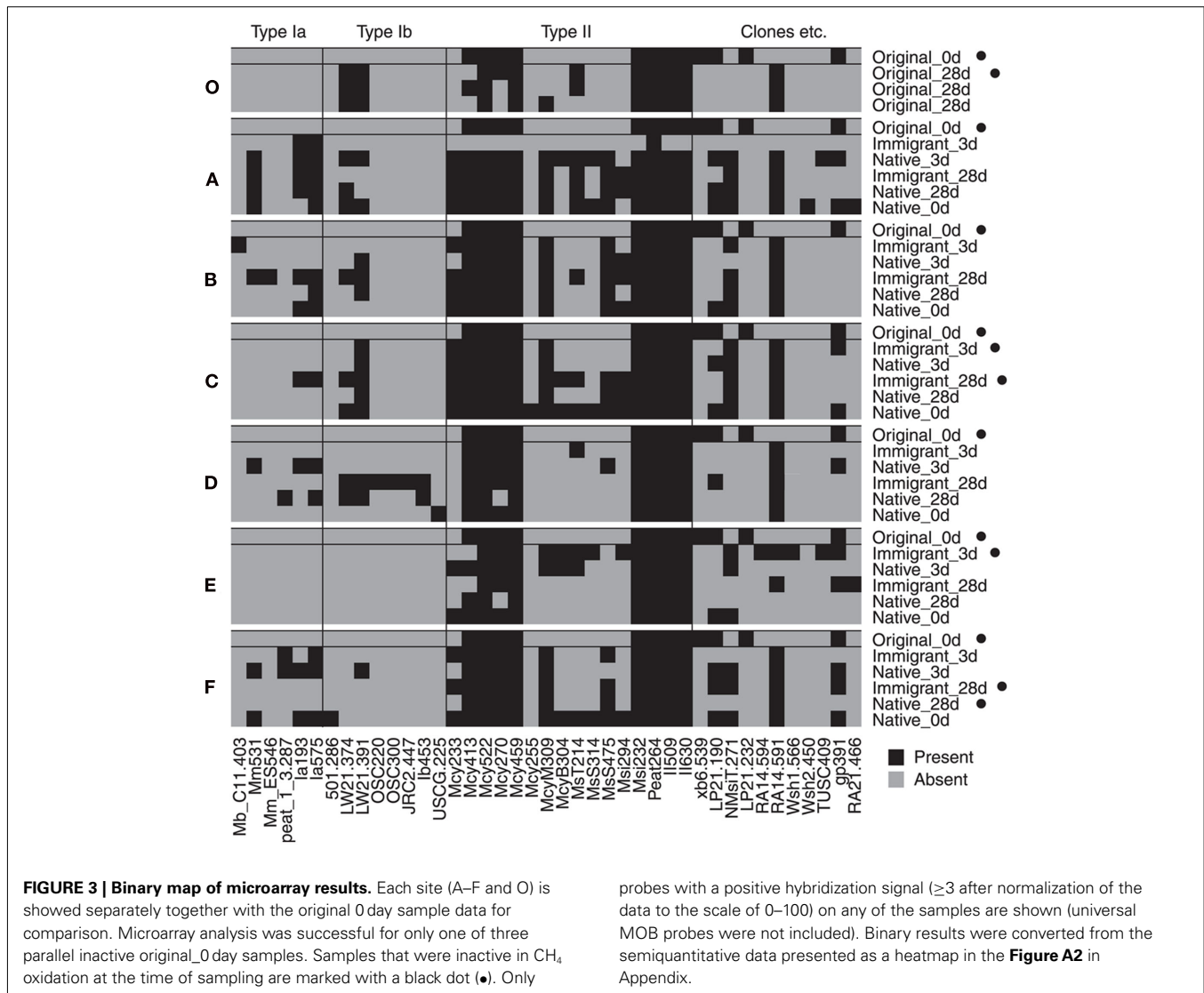
## RESULTS AND DISCUSSION

In the transplantation experiment SAM inactive *Sphagnum rubellum* mosses were planted on six different SAM active sites. Most of the originally inactive mosses showed detectable  $\text{CH}_4$  oxidation potentials ( $>0.005 \mu\text{mol CH}_4 \text{g}^{-1} \text{h}^{-1}$ ) after 3 days and 28 days (Figure 1). Comparing transplantation sites individually indicated that initially different MOB communities became more similar with time (Figure A1 in Appendix). Averaging over the entire data set, the *pmoA*-microarray showed that the MOB community of the originally inactive mosses started to resemble that of the native mosses in the active site after 3 days (Figure 2). After 28 days, MOB communities from the majority of inactive *Sphagnum* mosses transplanted to active sites (immigrants in Figure 1) were more similar to those of native active mosses than to those in the inactive site at 0 day. Thus, this field experiment indicated that MOB could be transferred between mosses through the water



**FIGURE 2 | Principal component analysis (PCA) of binary microarray data.** Sample means from different sampling times in days (d) are shown [original 0 days = original inactive moss ( $n = 1$ ), original 28 days = original moss transplanted back to the original site ( $n = 3$ ), immigrant = original moss transplanted to the active flark sites ( $n = 6$ ), native = native mosses of the active flark sites ( $n = 6$ )]. The first and second PCA axes explain 25 and 11% of the compositional variation, respectively.

phase. In addition, the original inactive *Sphagnum* (site O) became active after transplantation in its original site and had a different MOB community than before transplantation, demonstrating possible new MOB movement through the water phase (Figures 2 and 3 and Figure A1 in Appendix). However, although SAM activity was induced in most of the samples together with invasion of new MOB this was not always the case and SAM activity could also be induced without major changes in community composition (Site E, Figure A1 in Appendix). Thus there are methanotrophs that move through water but we cannot state that for all community members. Some members of the methanotroph community seem to be permanently associated with the mosses regardless of whether conditions favor  $\text{CH}_4$  oxidation or not. In our transplantation experiment this factor hindered us from seeing the invading members of the community and partly explains why large changes in community composition were not always seen when  $\text{CH}_4$  oxidation was reactivated. As *Methylocystis*-related type II MOB were present in practically all samples in the transplantation experiment (Figure 3 and Figure A2 in Appendix), as shown by the probes Mcy413, Mcy522, Mcy270, Mcy459, Msi232, Peat264, II509, and II630, their movement between mosses could not be clearly detected. Still, probes targeting *Methylocystis* strain M (McyM309), *Methylocystis* subgroup B (McyB304), and the genus *Methylosinus* (Mst214, Mst314, MsS475, Msi294, NMSiT.271) gave more variable presence patterns among the type II methanotrophs, suggesting water-mediated movement. Compared to  $\alpha$ -proteobacterial type II MOB, the other main methanotrophic



lineage, the  $\gamma$ -proteobacterial type I, was present more rarely and with more variability. Prior to transplantation, no type I methanotrophs were detected in inactive mosses. After 28 day, they were present in four of six “immigrant” samples transplanted to active sites. Type Ia subgroup was found only in mosses transplanted in the active flarks and the native mosses of these sites, suggesting that this group moved from the native mosses to the transplanted “immigrants.” Type I MOB could not be clearly linked to the emergence of CH<sub>4</sub> oxidation activity, although they were present in most of the active samples and absent from most of the inactive ones. Consequently, based on the transplantation experiment alone, we cannot exclusively state that invasion by nearby MOB is an imperative route in the reactivation of SAM activity.

To examine the hypothesis “the colonization of MOB through the water phase is a substantial reason for methanotrophic reactivation” further, a more simplified experiment was conducted in laboratory conditions. In this bathing experiment, SAM inactive *S. magellanicum* mosses exposed to unfiltered, SAM active flark water began CH<sub>4</sub> oxidation within 11 h, as

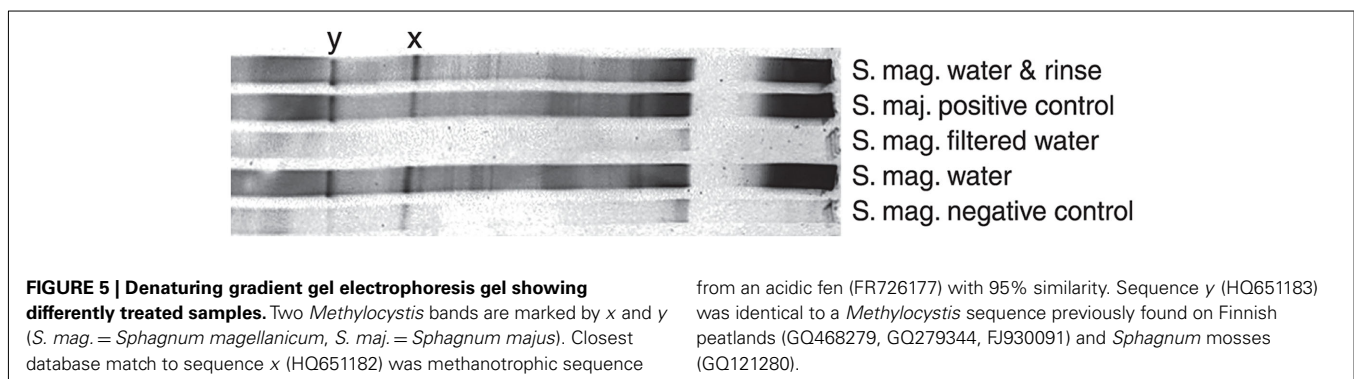
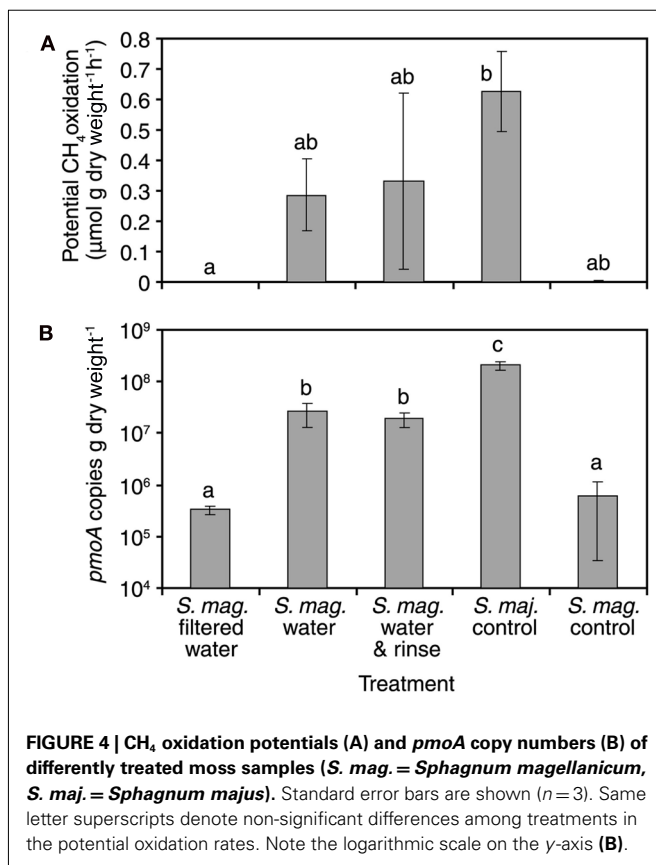
indicated by CH<sub>4</sub> oxidation potential measurements and community analyses including *pmoA*-based qPCR, DGGE fingerprinting and sequencing. Maximum CH<sub>4</sub> oxidation potentials and *pmoA* copy numbers were measured for mosses treated with unfiltered water (averages for rinsed mosses 0.33  $\mu\text{mol CH}_4 \text{g}^{-1} \text{h}^{-1}$ ,  $1.9 \times 10^7$  *pmoA* copies  $\text{g dw}^{-1}$ , and for unrinsed 0.29  $\mu\text{mol CH}_4 \text{g}^{-1} \text{h}^{-1}$ ,  $2.6 \times 10^7$  *pmoA* copies  $\text{g dw}^{-1}$ ) and positive control *S. majus* mosses (average 0.63  $\mu\text{mol CH}_4 \text{g}^{-1} \text{h}^{-1}$ ,  $20 \times 10^7$  *pmoA* copies  $\text{g dw}^{-1}$ ) from the active flark site (**Figure 4**). Respective values for mosses treated with filtered water (no CH<sub>4</sub> oxidation detected,  $0.03 \times 10^7$  *pmoA* copies  $\text{g dw}^{-1}$ ) and the negative control ( $<0.005 \mu\text{mol CH}_4 \text{g}^{-1} \text{h}^{-1}$ ,  $0.06 \times 10^7$  *pmoA* copies  $\text{g dw}^{-1}$ ) were clearly lower. The *pmoA* copy number between filtered and non-filtered flark water treated *S. magellanicum* increased in average by a factor of 63. Also, DGGE revealed the transfer of two *Methylocystis*-related methanotrophs through unfiltered water (**Figure 5**). Filtered water did not induce CH<sub>4</sub> oxidation activity. This experiment clearly demonstrated the water-mediated dispersal of MOB, but it also showed that compared to



invasion by new MOB, reactivation of the original MOB was not a major mechanism in the reactivation of CH<sub>4</sub> oxidation process in the studied mosses. In the DGGE gel the two *Methylocystis* bands are faintly present already in the unbathed negative control. Bathing of the mosses with sterile-filtered water caused these bands to fade away as shown also by the qPCR. On the contrary, treatment with unfiltered water caused emergence of high numbers of MOB and also high SAM activity. Thus the reactivation of the CH<sub>4</sub> oxidation activity must have been brought up by MOB invading the moss through the water phase or the growth in MOB numbers should have been seen also in the moss bathed with filtered water. Moreover, known *Methylocystis* strains have doubling times of several hours when growing on CH<sub>4</sub> in laboratory

conditions (Wise et al., 1999; Dedysh et al., 2007; Baani and Liesack, 2008) suggesting that it would be highly unlikely for these MOB to increase their numbers over 60 times higher in 11 h, as seen in the bathing experiment. Therefore we accept the posed hypothesis with a minor modification: MOB colonization through the water phase occurs and it obviously supports the reactivation of CH<sub>4</sub> oxidation in *Sphagnum* mosses, but our experiments cannot rule out the possibility of reactivation of original community members. This could be investigated in a prolonged bathing experiment in combination with diurnal light rhythm, but was beyond the scope of this investigation.

Based on our results that MOB colonize the mosses from water, the relationship between MOB and *Sphagnum* seems to be a loose, mutually beneficial association rather than a tight symbiosis. This result is in line with a recent finding by Bragina et al. (in press) who showed by pyrosequencing that some bacteria are passed from the *Sphagnum* sporophyte to the gametophyte but no known methanotrophs were among them. Representatives of the genus *Methylocystis*, however, were detected in the gametophyte. Still, even though methanotrophs may not be obligately dependent on the mosses, they most likely prefer the plant cells over life in the water phase. This is supported by our results from the bathing experiment. Despite having slightly lowered the amount of *pmoA* detected, mosses rinsed with sterile water had almost the same potential CH<sub>4</sub> oxidation activity as unrinsed ones, indicating that loosely attached methanotrophs play only a minor role in the process. In addition, the rapid (<11 h) increase in *pmoA* copy number suggests that methanotrophs present in the water phase quickly colonize *Sphagnum*. Compared to free-living bacteria, those associated with plants may gain an advantage from a stable CH<sub>4</sub> gradient and supply of oxygen from photosynthesis, but it has yet to be demonstrated. In the bathing experiment methanotrophs moved to the mosses even in the dark when oxygen was not formed in photosynthesis, indicating that at the very least oxygen is not the only advantage bacteria gain from the mosses. In another study no CH<sub>4</sub> oxidation activity was detected in peat water surrounding *Sphagnum* mosses (Kip et al., 2010), also indicating that, although present, MOB are not actively oxidizing CH<sub>4</sub> in the water phase. Since the only MOB, *Methylocystis*, detected in the bathing experiment, is non-motile (Dedysh, 2009), it remains open how these MOB cells end up on the moss surface and inside the hyaline cells. On the other hand another peatland inhabiting type II MOB genus, *Methylosinus* (Dedysh et al., 2003; Chen et al.,



2008), does contain motile species (Bowman et al., 1993) and has been isolated from *Sphagnum* mosses (Kip et al., 2011a).

Similar to previous studies (Tuomivirta et al., 2009; Larmola et al., 2010; Yrjälä et al., 2011) of Finnish peatlands, DGGE of our bathing experiment samples from Sallie's Fen, located in NH, USA, detected only *Methylocystis*-like MOB. These were also the dominant methanotrophs in our transplantation experiment, run on the Finnish Lakkasuo raised bog complex, when the *pmoA*-microarray was used. Dominance of type II MOB in our samples is in line with previous studies. Especially the high prevalence of *Methylocystis* in *Sphagnum* samples is not surprising as it is commonly found in northern peatlands (McDonald et al., 1996; Morris et al., 2002; Jaatinen et al., 2005; Dedysh et al., 2006; Chen et al., 2008). Although Kip et al. (2010) found, in contrast to our results, a high diversity of type I methanotrophs, *Methylocystis* was still the dominant species in their globally gathered *Sphagnum* samples (Kip et al., 2010) and also very abundant in mosses from a Dutch peat bog (Kip et al., 2011b).

We have demonstrated that water serves as an essential route for methanotroph dispersal and is thus an imperative part of *Sphagnum*-methanotroph association. This is likely to act as a backup mechanism for peatland CH<sub>4</sub> dynamics. Drainage of peatlands can alter the methanotroph community composition (Jaatinen et al., 2005; Yrjälä et al., 2011) and reduce *Sphagnum* coverage (Yrjälä et al., 2011), consequently compromising this mutualistic association. A case study (Yrjälä et al., 2011) found that the particular *Methylocystis* sp., which was found now also in the mosses of Sallie's Fen (northeastern USA) of our bathing experiment and in the mosses of Lakkasuo (Larmola et al. (2010)), was lost when the WT dropped by 14 cm, which is similar to the predicted drawdown for northern peatlands in the global warming scenario by 3°C

(Roulet et al., 1992). Restoration of drained peatlands aims to reactivate ecosystem function and restart methanogenesis (Tuittila et al., 2000). Any peatland restoration program should also aim to re-establish the conditions for the mutualistic association between methanotrophs and *Sphagnum*. Our study indicates that this could be done via transplantations of *Sphagnum* from donor sites with undisturbed CH<sub>4</sub> dynamics. In natural environments *Sphagnum* associated methanotrophic communities may reduce the methane flux by as much as 80% (Kip et al., 2010). It is not yet known whether this phenomenon can reach that scale also in compromised ecosystems.

## CONCLUSION

Here we showed, by two complementing experiments, that invasion of new MOB through water occurs and that it can be an important mechanism in the reactivation of CH<sub>4</sub> oxidation in *Sphagnum* mosses. Based on this result, the relationship between *Sphagnum* and methanotrophs is a loose, mutually beneficial association, although some methanotrophs may have an even tighter connection to the mosses.

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## APPENDIX

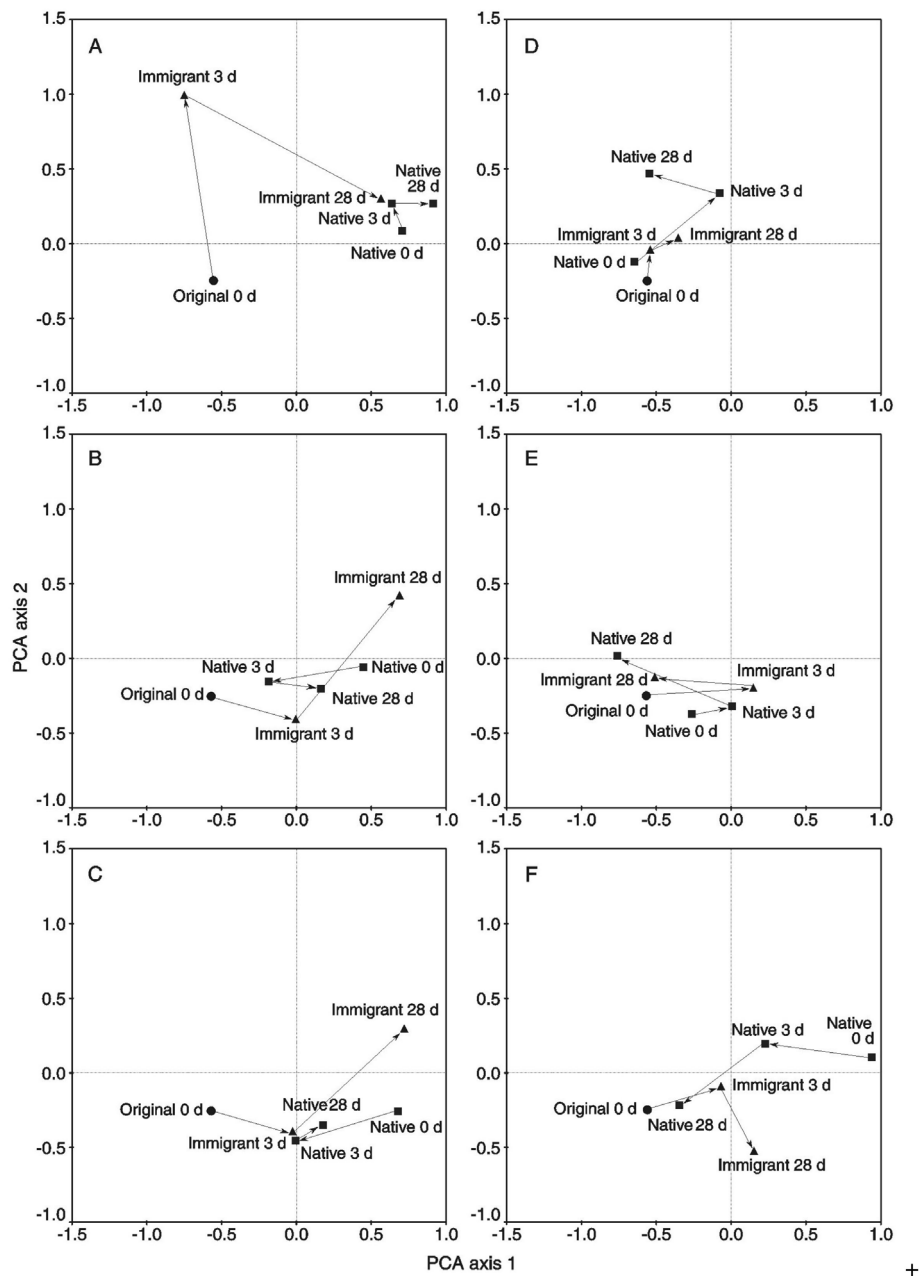
Table A1 | Ecological variables of Lakkasuo sites in the beginning (0 day) and the end (28 days) of the transplantation experiment.

Site	Sphagnum species	WT <sup>a</sup> (cm)		Peat temperature (°C)		CH <sub>4</sub> concentration in the pore water (ppm)	
		Start	End	Start	End	Start	End
A	<i>S. papillosum</i>	-5	-3	13.1	15.7	390	51.5
B	<i>S. papillosum</i>	-4	-8	13.8	15.7	3960	915
C	<i>S. papillosum</i>	-5	-6	13	15.1	5710	689
D	<i>S. cuspidatum</i>	-6	-5	12.1	14.8	397	201
E	<i>S. cuspidatum</i>	-7	-7	13	15.7	205	4.22
F	<i>S. majus</i>	-5	-5	13.3	15.4	3460	274
O	<i>S. rubellum</i>	-12	-22	15.0	15.3	16800	8740

<sup>a</sup>WT = water table level as measured from the peat surface.

Sites A–F = active sites where inactive original *Sphagna* were transplanted from site O.





**FIGURE A1 |** The principal component analysis (PCA) results of the binary microarray data divided into transplantation sites (A–F) (original 0 day = original inactive moss, original 28 days = original moss transplanted back to the original site,

immigrant = original moss transplanted to the active flark site, native = native moss of the active flark site). The first and second PCA axes explain 25 and 11% of the compositional variation, respectively.

